#### **FACT SHEET**

# REISSUANCE OF A GENERAL VPDES PERMIT FOR DISCHARGES FROM PETROLEUM CONTAMINATED SITES, GROUNDWATER REMEDIATION, AND HYDROSTATIC TESTS

The Virginia State Water Control Board has under consideration the reissuance of a general permit for discharges from petroleum contaminated sites, discharges from groundwater remediation, and discharges associated with hydrostatic testing. This general permit will replace the General VPDES Permit for Discharges from Petroleum Contaminated Sites and Hydrostatic Tests, VAG83, which expires February 25, 2008. Owners covered under the expiring general permit, who wish to continue to discharge under a general permit, must register for coverage under the new general permit.

Permit Number: VAG83

Name of Permittee: Any owner in the Commonwealth of Virginia agreeing to be regulated under the

terms of this general permit.

Facility Location: Commonwealth of Virginia

Receiving Waters: Surface waters within the boundaries of the Commonwealth of Virginia, except

waters specifically named in Board Regulations or Policies which prohibit such

discharges.

On the basis of preliminary review and application of lawful standards and regulations, the State Water Control Board proposes to issue the general permit subject to certain conditions and has prepared a draft permit. The Board has determined that this category of discharges is appropriately controlled under a general permit. The category of discharges to be included involves facilities with the same or similar types of operations and the facilities discharge the same or similar types of wastes. The draft general permit requires that all covered facilities meet standard effluent limitations, conditions and monitoring requirements.

Persons may comment in writing on the proposed issuance of the general permit within 60 days from August 20, 2007. Comments should be addressed to the contact person listed below. Comments shall include the name, address, and telephone number of the writer, and shall contain a complete, concise statement of the factual basis for comments. Only those comments received within this period will be considered by the Board.

All pertinent information is on file and may be inspected, and arrangements made for copying by contacting James Barnett at:

Virginia Department of Environmental Quality P.O. Box 1105 Richmond, Virginia 23218 (804) 698-4289 jsbarnett@deq.virginia.gov

A public hearing will be held on this draft permit. Notice of the public hearing will be published in newspapers and in the Virginia Register. Following the public hearing comment period, the Board will make its determinations regarding the proposed issuance.

# 1.0 Activities Covered By This General Permit

Petroleum contamination can occur as a result of leaks from above ground or underground storage tanks, pipeline leaks, surface oil spills and poor housekeeping at facilities that handle petroleum products. When the structural integrity of storage tanks or pipelines is tested with water pressure, the water may become contaminated with petroleum products. Chlorinated hydrocarbon solvents may be released into the environment via leakage from tanks, lines, process-related equipment, and spillage during materials handling operations. For the purposes of this general permit, "petroleum products" means petroleum-based substances comprised of a complex blend of hydrocarbons derived from crude oil such as motor fuels, jet

fuels, distillate fuel oils, residual fuel oils, lubricants, petroleum solvents and used oils. Petroleum products do not include hazardous waste as defined by the Virginia Hazardous Waste Regulations, 9 VAC 20-60. "Chlorinated hydrocarbon solvents" means solvents containing carbon, hydrogen, and chlorine atoms and the constituents resulting from the degradation of these chlorinated hydrocarbon solvents.

Contaminants may be introduced into surface waters when potable, or non-potable waters are used to hydrostatically test new or repaired petroleum or natural gas pipelines or petroleum storage tanks. These tests are commonly done in the pipeline industry and even though the events are usually sporadic in nature, they may produce a discharge significant in volume. Therefore, a general permit would adequately govern this type of activity.

This general permit will cover point source discharges of wastewaters from sites contaminated by petroleum products and chlorinated hydrocarbon solvents and also the point source discharges of hydrostatic test wastewaters resulting from the testing of petroleum and natural gas storage tanks and pipelines. These wastewaters may be discharged from the following activities: excavation dewatering, purging groundwater monitoring wells, conducting aquifer tests to characterize site conditions, hydrostatic tests of natural gas and petroleum storage tanks or pipelines, hydrostatic tests of underground and above ground storage tanks, pumping contaminated groundwater to remove free product from the ground, or discharges resulting from another petroleum product or chlorinated hydrocarbon solvent cleanup activity approved by the Department.

The effluent limits in the proposed general permit are established according to the type of petroleum product or chlorinated hydrocarbon solvent causing the contamination and the nature of the waterbody receiving the discharge. An option was being considered to allow less stringent effluent limits for small discharges that occur within a period of 72 consecutive hours, and with at least 3 years between occurrences of these discharges, but due to U.S. EPA concerns for impacts to threatened and endangered species, it was decided not to allow less stringent effluent limits for these discharges.

# 2.0 Proposed Effluent Limitations and Monitoring Requirements

# 2.1 Discharges of Water Contaminated with Gasoline

#### Freshwater Receiving Water Not Listed as Public Water Supply

Parameter	Limitation
Flow	No limit, monitoring required
Benzene	50.0 μg/l instantaneous max.
Toluene	175.0 μg/l instantaneous max.
Ethylbenzene	320.0 μg/l instantaneous max.
Total Xylenes	33.0 μg/l instantaneous max. e (1.273(ln hardness)) - 3.259
Total Recoverable Lead <sup>1</sup>	e (1.273(ln hardness)) - 3.259
Hardness <sup>1</sup>	mg/l, no limit
Ethylene Dibromide (EDB) <sup>1</sup>	5.3 μg/l instantaneous max.
1,2 Dichloroethane (1,2 DCA) <sup>1</sup>	990.0 μg/l instantaneous max.
pH	6.0 inst. min 9.0 inst. max.
MTBE	1,840.0 µg/l instantaneous max.
Ethanol <sup>2</sup>	4,100.0 µg/l instantaneous max.

<sup>&</sup>lt;sup>1</sup> Monitoring this parameter is required only when contamination results from leaded fuel. The minimum hardness concentration that will be used to determine the lead effluent limit is 25 mg/l.

The monitoring frequency is once per month. The permittee may request in writing that the monitoring frequency for ethanol be reduced to once per quarter if monitoring results from the first year of permit coverage demonstrate full compliance with the effluent limits.

<sup>&</sup>lt;sup>2</sup> Monitoring for ethanol is only required for discharges of water contaminated by gasoline containing greater than 10% ethanol.

Sites, Groundwater Remediation, and Hydrostatic Tests, VAG83

Freshwater Receiving Water Listed as a Public Water Supply

ParameterLimitationFlowNo limit, monitoring requiredBenzene12.0 μg/l instantaneous max.Toluene175.0 μg/l instantaneous max.

Ethylbenzene 320.0 μg/l instantaneous max. Total Xylenes 33.0 μg/l instantaneous max.

Total Recoverable Lead<sup>1</sup> Lower of: e (1.273(ln hardness)) - 3.259 or 15.0 µg/l

Hardness<sup>1</sup> mg/l, no limit

Ethylene Dibromide (EDB)<sup>1</sup>
1,2 Dichloroethane (1,2 DCA)<sup>1</sup>
3.8 μg/l instantaneous max.

pH
6.0 inst. min.- 9.0 inst. max.

MTBE
15.0 μg/l instantaneous max.

4,100.0 μg/l instantaneous max.

The monitoring frequency for all constituents or parameters is twice per month for the first year. If the first year's results demonstrate full compliance with the effluent limitations, the permittee may request that the monitoring frequency for ethanol be reduced to once per quarter and the other parameters to once per month.

# Saltwater Receiving Water body

Parameter	Limitation
Flow	No limit, monitoring required
Benzene	50.0 μg/l instantaneous max.
Toluene	500.0 μg/l instantaneous max.
Ethylbenzene	4.3 μg/l instantaneous max.
Total Xylenes	74.0 μg/l instantaneous max.
Total Recoverable Lead <sup>1</sup>	8.5 μg/l instantaneous max.
Ethylene Dibromide (EDB) <sup>1</sup>	5.3 μg/l instantaneous max.
1,2 Dichloroethane (1,2 DCA) <sup>1</sup>	990.0 µg/l instantaneous max.
рН	6.0 inst. min 9.0 inst. max.
MTBE	440.0 μg/l instantaneous max.
Ethanol <sup>2</sup>	4,100.0 μg/l instantaneous max.

<sup>&</sup>lt;sup>1</sup> Monitoring this parameter is required only when contamination results from leaded fuel. The minimum hardness concentration that will be used to determine the lead effluent limit is 25 mg/l. <sup>2</sup> Monitoring for ethanol is only required for discharges of water contaminated by gasoline

The monitoring frequency for all parameters and constituents is once per month. The permittee may request in writing that the monitoring frequency for ethanol be reduced to once per quarter if monitoring results from the first year of permit coverage demonstrate full compliance with the effluent limits.

# 2.2 Discharges of Water Contaminated with Petroleum Products Other than Gasoline

# Freshwater Receiving Water Not Listed as a Public Water Supply

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Parameter		Liı	mitation		

Flow No limit, monitoring required Naphthalene 10.0 µg/l instantaneous max.

<sup>&</sup>lt;sup>1</sup> Monitoring this parameter is required only when contamination results from leaded fuel. The minimum hardness concentration that will be used to determine the lead effluent limit is 25 mg/l.

<sup>&</sup>lt;sup>2</sup> Monitoring for ethanol is only required for discharges of water contaminated by gasoline containing greater than 10% ethanol.

containing greater than 10% ethanol.

Total Petroleum Hydrocarbons 15.0 mg/l instantaneous max.

The monitoring frequency for all parameters is once per month.

# Freshwater Receiving Water Listed as a Public Water Supply

rarameter	Limitation
Flow	No limit, monitoring required
Naphthalene	10.0 μg/l instantaneous max.
Total Petroleum Hydrocarbons	15.0 mg/l instantaneous max.
Н	6.0 inst. min 9.0 inst. max.
Benzene	12.0 μg/l instantaneous max.
MTBE	15.0 μg/l instantaneous max

The monitoring frequency for all parameters or constituents is twice per month for the first year. If the first year's results demonstrate full compliance with the effluent limitations, the permittee may request in writing that the monitoring frequency be reduced to once per month.

6.0 inst. min.- 9.0 inst. max.

# Saltwater Receiving Water body

рН

Parameter	Limitation
Flow	No limit, monitoring required
Naphthalene	8.9 μg/l instantaneous max.
Total Petroleum Hydrocarbons	15.0 mg/l instantaneous max.
pН	6.0 inst. min 9.0 inst. max.

The monitoring frequency for all parameters is once per month.

# 2.3 Discharges of Water from Hydrostatic Tests

# All Receiving Waters

Parameter	Limitation
Flow	No limit, monitoring required
pН	6.0 to 9.0 standard units
Total Petroleum Hydrocarbons (TPH)	15.0 mg/l instantaneous max.
Total Organic Carbon (TOC)	No limit, monitoring required
Total Suspended Solids (TSS)	No limit, monitoring required
Total residual chlorine (TRC)	.011 mg/l instantaneous max.

The monitoring frequency for all parameters is once per discharge.

# 2.4 Discharges of Water Contaminated by Chlorinated Hydrocarbon Solvents

# All Receiving Waters

Parameter	Limitation
Flow	No limit, monitoring required
chloroform	100.0 μg/l instantaneous max.
1,1 dichloroethane	4.0 μg/l instantaneous max.
1,2 dichloroethane	3.8 µg/l instantaneous max.
1,1 dichloroethylene	7.0 µg/l instantaneous max.
Cis 1,2 dichloroethylene	70.0 μg/l instantaneous max.
Trans 1,2 dichloroethylene	100.0 μg/l instantaneous max.
Methylene chloride	5.0 μg/l instantaneous max.
tetrachloroethylene	5.0 μg/l instantaneous max.
1,1,1 trichloroethane	112.0 μg/l instantaneous max.
1,1,2 trichloroethane	5.0 μg/l instantaneous max.
trichloroethylene	5.0 μg/l instantaneous max.

vinyl chloride2.0 μg/l instantaneous max.carbon tetrachloride2.5 μg/l instantaneous max.1,2 dichlorobenzene15.8 μg/l instantaneous max.chlorobenzene3.0 μg/l instantaneous max.trichlorofluoromethane5.0 μg/l instantaneous max.chloroethane3.6 μg/l instantaneous max.pH6.0 inst. min.- 9.0 inst. max.

The monitoring frequency for discharges into surface waters not listed as a public water supply (PWS) is once per month. The monitoring frequency for discharges into surface waters listed as a PWS is twice per month for the first year of permit coverage. If the permittee is in complete compliance with all effluent limitations, they may request that the monitoring frequency be reduced to once per month.

#### 3.0 Other Permit Conditions

The general permit prohibits discharge of floating solids or visible foam in other than trace amounts.

A condition is proposed in order to clarify the requirement for reporting of effluent monitoring results. Discharge monitoring is required each month in which a discharge occurs. For months when no discharge occurs, the permittee must submit a DMR certifying that there was no discharge. This system will allow DEQ to verify that either the effluent met the permit limits or that there was no discharge during the month.

Permittees that discharge treated wastewater are required to develop an Operations and Maintenance manual for the permitted treatment works. This requirement is imposed to assure proper operation and maintenance of facilities discharging under the general permit.

In order to assure that the proposed cleanup is conducted according to the methods in the approved Registration Statement, the permittee must construct treatment works prior to discharging and the permittee must notify the Department within 5 days of commencement of operation.

The general permit contains a condition designed to prevent pollution from materials stored on the site, which are not otherwise controlled by the effluent limitations.

If the proposed discharge is to surface waters via a municipal storm sewer system, the general permit requires the permittee to notify the owner of the storm sewer system. This is required in order to facilitate the municipality's efforts to control dry weather flows from the storm sewer.

A request for termination of coverage under the permit is required to provide documentation for the permittee and the Department that the activities covered under the general permit have been concluded and coverage is no longer needed.

The general permit anticipates that the covered treatment works will not be treating sewage from other users or indirect dischargers. Therefore, the permit contains no conditions applicable to such users. This permit also does not cover treated sewage discharges from the permittee or other users.

# 4.0 Discharges to Public Water Supplies (PWS)

This permit may be used to authorize discharges to a PWS. The Virginia Department of Health, Office of Water Supply Programs generally requires a minimum of 5 miles separation between a discharge and a PWS intake (12 VAC 5-590-200). This general permit will use the same separation distance. Discharges into a surface water designated as a PWS will not be allowed under this permit if the discharge location is less than 5 miles upstream of the PWS intake.

# 5.0 Revisions to Expiring VPDES General Permit for Petroleum Contaminated Sites and Hydrostatic Tests

The proposed regulation allows discharges to waters designated as a PWS as long as the discharge location is at least five miles upstream of the PWS intake.

The proposed regulation has been expanded in scope over the present and previous versions of this regulation and may be used to permit discharges of water contaminated by chlorinated hydrocarbon solvents.

Effluent limits for dissolved lead, xylenes, and naphthalene have been revised. Effluent limits for ethylene dibromide, 1,2 dichloroethane, and ethanol have been added to this general permit. The basis for each of these changes is discussed below.

#### 6.0 Basis for Effluent Limitations

### 6.1 Discharges of Gasoline Contaminated Water

This general permit contains both technology-based and water quality-based effluent limits. Where both types of limits were available, the more stringent of the two was chosen. The U.S. EPA has developed a model NPDES permit for discharges from gasoline contaminated underground storage tank sites. The model permit provides technology-based effluent limitations for surface water discharges. The technology basis for those limitations is free product removal followed by air stripping. The limits are set for benzene and the sum of benzene, toluene, ethylbenzene, and xylenes (BTEX). These parameters are used as indicators of the compounds most likely to be found in gasoline. Benzene is considered a good indicator of the removal of volatile organic gasoline constituents via air stripping because of its relatively high water solubility and low volatility compared to other gasoline components.

The EPA model permit states that air strippers have the potential to operate at 99.5% efficiency and it uses this as the basis for limitations on benzene and BTEX. However, it also states that one cannot assume optimal operational conditions at all times and that permit limitations must be achievable with existing technology at reasonable cost. The model permit then establishes optional limitations based on 95% removal efficiency. The 95 percent efficiency rating accounts for operational difficulties which may be encountered during periods of low temperature and/or high humidity when air strippers may not be expected to perform at the 99.5% peak efficiency level. The EPA Treatability Database (RREL Version 5.0) contains information on treatment of the BTEX compounds at various concentrations by air stripping and granular activated carbon. The average removal efficiencies in contaminated groundwater are as follows: benzene 97%, toluene 97.4%, ethylbenzene 87% and xylene 88%. The 95% removal efficiency also provides the possibility for considerable cost savings for the tank owners/operators involved in cleaning up underground storage tank (UST) sites, many of whom are small businesses without the resources to install state-of-the-art equipment. The number of sites cleaned up under the Virginia Petroleum Storage Tank Fund would also increase if the cost per site were less.

The technology-based benzene limit of  $50~\mu g/l$  in the EPA model permit is derived by assuming a concentration of 1 mg/l benzene in the influent to the treatment system and 95% removal. Thus, the technology-based limitations of  $50~\mu g/l$  in this general permit are based on the 95% removal efficiency assumption allowed in the EPA model permit.

The water quality-based effluent limitations in this general permit are established pursuant to the VPDES Permit Regulation, 9 VAC 25-31-220 D, and the policy stated in the Virginia Water Quality Standards, 9 VAC 25-260-140 B. The limits are set at what are believed to be safe concentrations for the protection of beneficial uses including the growth and propagation of aquatic organisms inhabiting surface waters which receive the discharge. They assume zero dilution of the effluent by the receiving waters so that they can be applied without regard to effluent or receiving water flows. They are based on information provided in EPA criteria documents for priority pollutants, EPA toxicity databases and conservative application factors.

The aggregate parameter BTEX is used in the EPA model NPDES permit previously discussed to limit 4 parameters. It sets an effluent limitation for BTEX at 750  $\mu$ g/l based on an assumed influent BTEX concentration of 15 mg/l and the 95% air stripper removal efficiency. The model permit document states that the composition of gasoline is highly variable and any one of the four BTEX components may be the primary constituent. The discussion of water quality-based limits which follows identifies cases where the 750  $\mu$ g/l technology-based limitation on BTEX would not protect aquatic life from adverse effects.

In some circumstances, if a specific BTEX component were to dominate the mixture the resulting effluent could be toxic at, or below 750  $\mu$ g/l. For instance, Thomas and Delfino (1991) found that toluene comprises about 50% of the total BTEX in gasoline when analyzed by EPA Methods 610 and 602. If the BTEX limit were set at 750  $\mu$ g/l then this could allow up to 375  $\mu$ g/l of toluene in an effluent. The discussion on water quality-based limits which follows sets a limit of 175  $\mu$ g/l for toluene in discharges to freshwater. The same researchers found that xylenes made up about 30% of the total BTEX in gasoline. When applied to the 750  $\mu$ g/l BTEX limit in the EPA model permit this results in a possible xylene discharge level of 225  $\mu$ g/l. Based on available information, total xylenes should not exceed 33  $\mu$ g/l in freshwater. Without limits on individual parameters, ethylbenzene in discharges to saltwater could still be chronically toxic at the 100  $\mu$ g/l BTEX technology-based limit given in the model permit using 99.5% removal efficiency.

Based on this discussion, the general permit does not contain a technology-based BTEX limit. Instead, it establishes water quality-based limits on the individual components (benzene, toluene, ethylbenzene and total xylenes), which result in lower total BTEX levels in the discharge. When the proposed limits for individual components are summed, the BTEX value for the freshwater discharges is  $627 \mu g/l$  and for discharges to saltwater the value is  $628.3 \mu g/l$ .

#### 6.1.1 Benzene

#### **Freshwater**

The EPA criteria document for benzene (EPA 440/5-80-018, EPA 1980a) states that benzene may be acutely toxic to freshwater organisms at concentrations as low as 5,300  $\mu$ g/l. This is an LC50 value for rainbow trout. The document also states that acute toxicity would occur at lower concentrations among more sensitive species. No data were available concerning the chronic toxicity of benzene to sensitive freshwater organisms. The derivation of a "safe level" for benzene was based on the 5,300  $\mu$ g/l LC50. This value was divided by 10 in order to approximate a level which would not be expected to cause acute toxicity. (The use of an application factor of 10 was recommended by the National Academy of Sciences in the EPA's publication "Water Quality Criteria, 1972" (EPA/R3/73-033). This use of application factors when setting water quality criteria is still considered valid in situations where data are not sufficient to develop criteria according to more recent guidance.) The resulting "non-lethal" concentration of 530  $\mu$ g/l was divided by an assumed acute to chronic ratio of 10 to arrive at the water quality-based permit limitation of 53  $\mu$ g/l. (When actual data are not available, EPA, in the Technical Support Document for Water Quality-based Toxics Control (EPA/505/2-90-001) recommends using an acute to chronic ratio of 10). The EPA model permit's technology-based 50  $\mu$ g/l value for benzene is more protective, therefore, it was chosen over the 53  $\mu$ g/l water quality-based concentration.

The Virginia Water Quality Standard Regulation (9 VAC 25-260) contains a human health standard of 710  $\mu$ g/l for benzene in surface waters that are not a PWS. This concentration is well above the aquatic toxicity concentration of 53  $\mu$ g/l and the technology-based concentration of 50  $\mu$ g/l. The lowest of these concentrations is the technology-based limit of 50  $\mu$ g/l and it is recommended that the effluent limit for benzene in freshwaters that are not a PWS be set at 50  $\mu$ g/l.

## Saltwater

The limited data for benzene and saltwater organisms in the EPA criteria document indicates that stress and survival effects occur at concentrations as low as 700  $\mu$ g/l when fish are exposed for long periods. Based on the application of a 0.10 safety factor to this chronic effect concentration, the water quality-based limit for discharges to saltwater would be 70  $\mu$ g/l. Once again, the 50  $\mu$ g/l technology-based limitation is recommended because it is attainable and more protective.

The Virginia Water Quality Standard Regulation (9 VAC 25-260) contains a human health standard of 710 µg/l for benzene in surface waters that are not a PWS. This concentration is well above the saltwater organism, chronic toxicity concentration of 70 µg/l and the technology-based concentration of 50 µg/l. The lowest of these concentrations is the technology-based limit of 50 µg/l and it is recommended that the

effluent limit for benzene in saltwater be set at 50 µg/l.

# **Public Water Supplies**

The Virginia Water Quality Standard Regulation (9 VAC 25-260) contains a human health standard of 12  $\mu$ g/l for benzene in public water supplies. This concentration is well below the aquatic toxicity concentration of 53  $\mu$ g/l and the technology-based concentration of 50  $\mu$ g/l. The human health standard of 12  $\mu$ g/l is recommended as the effluent limit for benzene in public water supplies.

# 6.1.2 Ethylbenzene

#### **Freshwater**

The EPA criteria document for ethylbenzene (EPA 440/5-80-048, 1980b) gives an acute effects concentration of 32,000  $\mu$ g/l. This is an LC50 for bluegill sunfish. Acute toxicity may occur at lower concentrations if more sensitive species were tested. No definitive data are available on the chronic toxicity of ethylbenzene to freshwater organisms. In order to derive an acceptable level of ethylbenzene for the protection of freshwater organisms the acute value of 32,000  $\mu$ g/l was divided by 100, using the same assumptions employed above for benzene. The resulting value of 320  $\mu$ g/l is a calculated chronic toxicity concentration for ethylbenzene.

The human health water quality standard for ethylbenzene in surface waters that are not a PWS is 29,000  $\mu$ g/l. The chronic toxicity concentration of 320  $\mu$ g/l is well below the human health standard and is the recommended effluent limit.

#### Saltwater

According to the criteria document, ethylbenzene is acutely toxic to certain saltwater organisms at concentrations as low as 430  $\mu$ g/l and may be acutely toxic at lower concentrations if more sensitive organisms are tested. Dividing this number by the 100 application factor yields the proposed effluent limit of 4.3  $\mu$ g/l for discharges to saltwater receiving waters.

#### **Public Water Supplies**

The Virginia human-health water quality standard for ethylbenzene in public water supplies is  $3,100 \mu g/l$ . The freshwater effluent limit based on aquatic toxicity is more stringent than human-health based standard for public water supplies and should be protective of human health concerns.

#### 6.1.3 Toluene

The EPA criteria document for toluene (EPA 440/5-80-075, 1980c) states that acute toxicity to freshwater organisms occurs at 17,500  $\mu$ g/l and would occur at lower concentrations if more sensitive organisms were tested. No data are available on the chronic toxicity of toluene to freshwater species. Based on the available data for acute toxicity and dividing by the application factor of 100, the proposed effluent limit for toluene discharged to freshwater is 175  $\mu$ g/l.

The available data indicate that toluene is chronically toxic to certain saltwater organisms at concentrations as low as  $5,000 \,\mu\text{g/l}$ . Chronic toxicity levels are expected to occur at lower concentrations if more sensitive organisms are tested. Dividing this chronic effects level by 10 resulted in the proposed saltwater discharge effluent limit of  $500 \,\mu\text{g/l}$ .

The Virginia human health standards for toluene in drinking and non-drinking water streams are 6,800  $\mu$ g/l and 200,000  $\mu$ g/l, respectively. The proposed effluent limits based on aquatic toxicity are more stringent than human health based standards and should be protective of human health. For discharges into public water supplies, it is recommended that the freshwater aquatic toxicity value of 175  $\mu$ g/l be used as the effluent limit.

#### 6.1.4 Xylenes

Xylene is not a 307(a) priority pollutant, therefore no criteria document exists for this compound. There are

three isomers of xylene (ortho, meta and para) and the general permit limits are established so that the sum of all xylenes is considered in evaluating compliance. The proposed effluent limits are based on a search of the EPA's ECOTOX data base. According to ECOTOX, the lowest freshwater LC50 for xylenes is 3,300  $\mu$ g/l reported for rainbow trout (Mayer and Ellersieck, 1986). Based on the rationale presented earlier for other compounds, this acutely toxic concentration was divided by 10 to account for species that were not tested but which may be more sensitive than rainbow trout. Then, in order to find a concentration that is expected to be safe over chronic exposures, an additional safety factor of 10 was applied to arrive at the proposed effluent limitation of 33  $\mu$ g/l total xylenes.

The LC50 of 7,400  $\mu$ g/l for grass shrimp (Neff et al., 1979) is the lowest saltwater value in the ECOTOX database. This LC50 concentration was divided by 100 to derive the saltwater effluent limit of 74  $\mu$ g/l total xylenes.

There is no Virginia human health water quality standard for xylenes. The Maximum Contaminant Level and Maximum Contaminant Level Goal for xylenes in the EPA Safe Drinking Water Regulation, 40 CFR Part 141, are both set at 10 mg/l ( $10,000 \mu g/l$ ). The proposed permit limits based upon aquatic toxicity are more stringent than drinking water standards for xylenes and are expected to be protective of human health.

#### 6.1.5 Lead

The EPA permit model for discharges of petroleum contaminated water does not contain a recommended effluent limit for lead. It is recognized that tetraethyl and tetramethyl lead may be present in gasoline at leaking storage tank sites. These organic lead compounds, if present, are expected to be removed via air stripping along with other volatile organics.

The proposed effluent limits for lead are based upon the Virginia Water Quality Standards for the protection of fresh and saltwater organisms to chronic exposure to lead. The effluent limit for lead in wastewater discharged into streams listed as public water supplies also must meet the water quality standard for lead in public water supplies. While the water quality standards require analysis for dissolved metals, this permit requires that samples be analyzed for Total Recoverable Lead as required by the Virginia Pollutant Discharge Elimination System (VPDES) Permit regulation 9 VAC 25-31-230C. The chronic standard for lead in saltwater when the general permit regulation was initially adopted was 8.5 µg/l. Less stringent water quality criteria were adopted by the Board on September 25, 1997. The lead standard for saltwater used in the existing general permit, however, cannot be revised due to anti-backsliding requirements and the effluent limit for lead discharged into saltwater must remain at 8.5 µg/l.

Virginia's freshwater lead standard for the chronic exposure of organisms to this constituent is based upon the hardness of the water in the waste stream. The lead standard for chronic toxicity to freshwater aquatic organisms is now calculated by equation (1) (Virginia Water Quality Standard Regulation, adopted September 25, 1997). The freshwater lead standard in the present general permit is more stringent than the lead standard in the 1997 Water Quality Standard Regulation and is calculated from equation (2). Equation (2) was taken from the freshwater lead standard for chronic toxicity listed in Virginia's 1992 Water Quality Standard Regulation (VR 680-21-00).

- (1)  $e^{(1.273(\ln \text{hardness})) 3.259}$
- (2) e (1.273(ln hardness)) 4.705

The proposed reissuance shall use equation (1) to calculate the aquatic toxicity-based lead effluent limit. The minimum hardness to be used in the calculation of the lead effluent limit is 25 mg/l. The change proposed with this reissuance conforms to the anti-backsliding provisions of Section 402(o) of the Clean Water Act, 9 VAC 25-31-220.L, and 40 CFR § 122.44. The limits proposed for lead are water quality based effluent limits. The revisions to the limits are allowed since the revisions comply with the water quality standards 402(o)(3) and they are consistent with antidegradation 303(d)(4)(B).

The Human Health water quality standard for lead in public water supplies is 15  $\mu$ g/l. When wastewater is discharged to a public water supply, the effluent limit will be the lower of 15  $\mu$ g/l or the calculated

aquatic toxicity based limit.

# 6.1.6 Ethylene Dibromide (EDB)

Ethylene dibromide (a.k.a. 1,2 dibromoethane, CAS Number: 106-93-4) is a compound added to leaded gasolines to remove lead from the combustion chamber and prevent lead oxide and lead sulfide deposits from forming within an internal combustion engine. Lead scavengers such as ethylene dibromide (EDB) are persistent in groundwater and, in combination with the BTEX constituents can be good indicators of a leaded gasoline release.

EPA has no criteria documents for EDB nor are there existing water quality standards for this constituent. According to the ECOTOX database, the lowest freshwater LC50 concentration for this constituent is  $15,000~\mu g/l$  for largemouth bass (Davis and Hardcastle, 1959). Dividing this LC50 value by 100 leads to a concentration of  $150~\mu g/l$ . In saltwater, the lowest LC50 is  $4800~\mu g/l$  for the sheepshead minnow (Landau and Tucker, 1984). Dividing this LC50 value by 100 leads to a saltwater aquatic toxicity value of  $48~\mu g/l$ .

The procedure used by Virginia for calculating water quality standards for human health involves using risk factors, average adult body weight, intake of water and fish (public water supplies) and fish only, and a bioconcentration factor for the constituent. Ethylene dibromide is considered a human carcinogen and equation (3) listed below is used by Virginia to derive human-health based water quality criteria for waters that are not public water supplies. Based upon an excess lifetime cancer risk of one in one hundred thousand and an oral carcinogenic potency slope factor of 2 mg/kg/day (EPA IRIS database, 2007c), a human health concentration of 5.3 µg/l was derived for EDB in surface waters that are not public water supplies. This human health concentration is much more stringent than the fresh or saltwater toxicity values and it is the recommended effluent limit for EDB in waters that are not listed as a PWS.

The federal drinking water standard for EDB is .05  $\mu$ g/l. Equation (4) shown below is used by Virginia to develop human health based water quality criteria for surface waters listed as public water supplies. Based upon an excess lifetime cancer risk of one in one hundred thousand and an oral carcinogenic potency slope factor of 2 mg/kg/day (EPA IRIS database, 2007c), a human health concentration of .169  $\mu$ g/l was derived for EDB in surface waters that are public water supplies. This human health concentration is the recommended effluent limit for EDB in surface waters listed as a PWS.

Equation to derive human health criteria for surface waters that are not a PWS:

Equation to derive human health criteria for public water supplies:

Where: Risk = excess lifetime cancer risk. The Water Quality Standards are based on an excess lifetime cancer risk of one in one hundred thousand risk level or 10<sup>-5</sup>

Adult body weight = 70 kg

CSFo = carcinogenic slope factor, oral exposure route (mg/kg/day)

Water intake = typical daily water intake for an adult, 2 1/day

FI = fish intake. The Water Quality Standards are based on a fish intake of .0065 kg/day BCF = bioaccumulation factor (l/kg)

Derivation of Human Health concentration for EDB in surface waters that are not a PWS:

$$1 \times 10^{-5} * 70 \text{ kg}$$
  
WQS = -----

$$2 \text{ mg/kg/day} * .0065 \text{ kg/day} * 10.2 \text{ l/kg}$$
  
WQS =  $5.3 \times 10^{-3} \text{ mg/l}$  or  $5.3 \mu \text{g/l}$ 

According to EXTOXNET DATABASE (1996), the bioaccumulation factor for EDB is 10.2 l/kg. The carcinogenic slope factor, oral exposure route for EDB is 2 mg/kg/day (EPA IRIS database, 2007c).

Derivation of Human Health concentration for EDB in surface waters that are a PWS:

$$WQS = \frac{1 \times 10^{-5} * 70 \text{ kg}}{2 \text{ mg/kg/day} * [2 \text{ l/day} + (.0065 \text{ kg/day} * 10.2 \text{ l/kg})]}$$
 
$$WQS = 1.69 \times 10^{-4} \text{ mg/l or } .169 \text{ µg/l}$$

# **6.1.7 1,2-Dichloroethane (1,2 DCA)**

Another compound commonly added to leaded gasoline as a lead scavenger is 1,2-Dichloroethane (1,2 DCA, CAS Number: 107-06-20). The EPA criteria document for chlorinated ethanes (EPA 440/5-80-029, 1980d) states that acute toxicity to freshwater organisms exposed to 1,2 DCA occurs at 118,000 µg/l and would occur at lower concentrations if more sensitive organisms were tested. No data are available on the chronic toxicity of 1,2 DCA to freshwater species. Based on the available data for acute toxicity and dividing by the application factor of 100, an aquatic toxicity limit for 1,2 DCA in freshwater is 1,180 µg/l.

The available data indicate that 1,2 DCA is acutely toxic to certain saltwater organisms at concentrations as low as 113,000  $\mu$ g/l. Based on the available data for acute toxicity and dividing by the application factor of 100, the aquatic toxicity limit for 1,2 DCA in saltwater is 1,130  $\mu$ g/l.

The Virginia human health standards for 1,2 DCA in surface waters that are public water supplies and surface waters that are not public water supplies are 3.8  $\mu$ g/l and 990  $\mu$ g/l, respectively. The human health criteria are more stringent than the aquatic toxicity criteria. It is recommended that a limit of 990  $\mu$ g/l be used for discharges to surface waters that are not public water supplies. For discharges into public water supplies, it is recommended that the Virginia public water quality criteria of 3.8  $\mu$ g/l be used.

# 6.1.8 Methyl Tertiary Butyl Ether

Methyl-tertiary-butyl ether (MTBE) is a common additive in "reformulated" automotive gasolines. This oxygenate is supposed to reduce winter-time carbon monoxide levels in U.S. cities. It also is believed to be effective in reducing ozone and other toxics in the air year-round. If MTBE is used, it can be present in gasoline at up to 15% of the volume of the fuel. MTBE is an extremely hydrophilic compound. Unlike most petroleum products, it readily dissolves in water. The presence of MTBE in gasoline can increase the solubility of the fuel mixture in groundwater. MTBE may be removed from contaminated groundwater by air stripping treatment technologies. However, due to its hydrophilic nature, a higher air/water ratio is required to remove this constituent via air stripping than is required for BTEX removal. According to the EPA Treatability Database (RREL Version 5.0), MTBE removal efficiency via air stripping ranges from approximately 63 percent to 79 percent. If the MTBE concentration in the system influent is 10 mg/l and removal efficiency of 75 percent is achieved, air stripping should be capable of reducing the MTBE concentration to 2.5 mg/l.

Neither EPA nor the DEQ has established water quality criteria for MTBE for protection of aquatic life or human health. Literature searches indicated several studies that evaluated the effects of MTBE on aquatic organisms. According to BenKinney et al. (1994), MTBE was acutely toxic (LC50) to green algae (*Selanastrum capricornutum*) at a concentration of 184,000  $\mu$ g/l. Geiger and associates (1988) found that MTBE was acutely toxic to the fathead minnow (*Pimephales promelas*) at a concentration of 672 mg/l (672,000  $\mu$ g/l). Application of the customary safety factor of 100 to the LC50 concentration for green algae results in a concentration of 1,840  $\mu$ g/l. This concentration is recommended as the discharge limit for MTBE into freshwater.

The literature search revealed several studies performed on the toxicity of MTBE to marine organisms. BenKinney et al. (1994) found that MTBE was acutely toxic to the inland silverside (*Menidia beryllinia*) at a concentration of 574 mg/l. According to Boeri and associates (1994), MTBE was acutely toxic to mysid shrimp (*Mysidopsis bahia*) at 44 mg/l (44,000  $\mu$ g/l). Application of the customary safety factor of 100 to the LC50 for the mysid shrimp results in a concentration of 440  $\mu$ g/l. A concentration of 440  $\mu$ g/l is recommended as the effluent limit for MTBE discharged into saltwater.

According to Fujiwara et al. (1984) and the European Fuel Oxygenates Association, bioaccumulation factors for MTBE in fish tissue are 1.5 l/kg and 1.6 l/kg, respectively. Moreover, Fujiwara found that discontinued exposure of the fish to MTBE caused fish to quickly excrete the MTBE remaining in their tissues.

Derivation of Human Health concentration for MTBE in surface waters that are not a PWS:

$$WQS = \frac{1 \times 10^{-5} * 70 \text{ kg}}{4 \times 10^{-3} \text{ mg/kg/day} * .0065 \text{ kg/day} * 1.6 \text{ l/kg}}$$

$$WQS = 16.827 \text{ mg/l} \text{ or } 16.827 \text{ µg/l}$$

NOTE: The Carcinogenic Slope Factor, oral exposure route of 4 X 10<sup>-3</sup> mg/kg/day is a value from the EPA Region III October 2006 Risk Based Concentration Table (EPA Region III, 2006).

The Virginia Department of Health, Office of Water Programs has established a trigger level of 15  $\mu$ g/l for MTBE in public drinking water. The U.S. EPA has established a drinking water health advisory for MTBE of 20 – 40  $\mu$ g/l based upon taste and odor effects. These levels are lower than the lowest concentration that caused observable effects in animals. For waters designated as a PWS, an effluent limit of 15  $\mu$ g/l for MTBE is recommended.

#### 6.1.9 Ethanol

Ethanol has been used in U.S. automotive gasolines for over thirty years. During the oil embargo of 1973, ethanol was used as a gasoline extender to counteract rising fuel prices and increase the nation's gasoline supply (Texas State Energy Conservation Office, 2007a). As lead was phased out of gasoline, ethanol and MTBE were used as octane enhancers in lieu of tetraethyl lead. Later, MTBE and ethanol were the primary products used to meet the standards for the Wintertime Oxygenated Fuels Program (1992) and Phase 1 and Phase 2 of the Reformulated Gasoline Program (RFG, 1995 and 2000). Ethanol was used primarily in gasoline sold in the Midwest and MTBE was used in gasoline sold in most of the rest of the U.S.

The federal Energy Policy Act of 2005 removed the oxygenate mandate for RFG and established a national renewable fuel standard (RFS; Meyers, 2006). Consequently, suppliers requested major pipelines to remove MTBE from RFG. In February 2006, Colonial Pipeline, which serves Virginia, announced that it would discontinue shipping RFG with MTBE (O'Connor, 2006). In the Spring of 2006, many RFG marketers in Virginia began being supplied with gasoline containing up to 10% ethanol (E10) in order to replace the MTBE.

The fate and transport of ethanol in groundwater is controlled primarily by biodegradation (Ulrich, 1999). Based on the chemical behavior of ethanol, it is expected that ethanol in subsurface releases of oxygenated gasolines will rapidly partition into groundwater and will become the dominant dissolved contaminant immediately downgradient of the release. It is believed that mechanisms for attenuating subsurface contaminants, such as sorption, volatilization, and abiotic degradation, will not substantially contribute to the decreased mobility or loss of ethanol in subsurface aquifers.

According to EPA (2000), ethanol is not expected to persist in the groundwater because it biodegrades readily nor does ethanol appear to pose as great a danger to groundwater supplies as does MTBE. Ethanol is considerably less volatile than MTBE in surface waters because it has a lower Henry's law constant (Layton and Daniels, 1999). Though ethanol's volatilization-loss rate from water is much less

than that of MTBE, ethanol will not persist in water because it undergoes fairly rapid biodegradation. Thus, ethanol is a short-lived compound in surface waters and subsurface aquifers.

Under the Clean Water Act, the EPA promulgated effluent limitations and standards controlling discharges from the production of organic chemicals, plastics, and synthetic fibers (EPA, 2005 and 2007a), and from pharmaceutical facilities with operations in fermentation; extraction; chemical synthesis; mixing, compounding, and formulating; and research (EPA, 1999 and 2007b). For certain pharmaceutical facilities directly discharging ethanol, the maximum daily discharge limit for ethanol is 10.0 mg/L, and the average monthly discharge must not exceed 4.1 mg/L.

Jack Hwang of EPA Region 3 performed initial research on discharge limits and extra parameters for monitoring blended fuel releases in response to inquiries from the State of Maryland and the Commonwealth of Virginia (Hwang, 2007). Based on discussions with an EPA regional toxicologist and with Dr. John Wilson, one of EPA's microbiologists, Mr. Hwang indicates that:

"There is no concern for human health risk - the limit would be very high, nor is there concern for toxicity to aquatic organisms. If there is a need for setting an ethanol limit, the most likely reason would be due to the consideration of "oxygen depletion" in surface water. However, the limit could be site specific depending on the characteristics of the receiving water body and the allowable dilution ratio."

Ethanol is a short-lived compound in the environment due to the ubiquity of microorganisms capable of metabolizing ethanol and to the rapid rates of ethanol biodegradation (Ulrich, 1999). Since ethanol is rapidly metabolized, it is unlikely that ethanol will travel a substantial distance once released into the subsurface or that it will persist in the subsurface or surface waters. It should be noted, however, for E85 (ethanol comprises 85% of the gasoline) releases or neat ethanol releases into surface waters, microorganisms involved with breaking down the ethanol could scavenge the available oxygen thereby creating anaerobic conditions and causing a fish kill (Kuhn, 2007). The same would likely hold true for large E10 releases into surface waters.

Neither the DEQ nor EPA has promulgated acute and chronic water quality criteria for ethanol in surface waters. Acute and chronic water quality benchmarks for ethanol were developed using toxicity information available for aquatic invertebrates (*Daphnia* species), rainbow trout, and the fathead minnow from EPA's ECOTOX database (Iott, 2001). Based on the available data and using Tier II procedures outlined in the for EPA's Final Water Quality Guidance for the Great Lakes System, an acute water quality benchmark for ethanol in surface water is 564 mg/L, and a chronic water quality benchmark for ethanol is 63 mg/L. The values indicate that an ethanol concentration of 564 mg/L in the water column is likely to cause acute toxicity to freshwater aquatic life and that an ethanol concentration of 64 mg/L in the water column is likely to cause chronic toxicity to freshwater life. The chronic and acute water quality benchmarks developed for ethanol (EPA, 2006) are lower than draft water quality criteria developed by the EPA.

The DEQ has limited experience in dealing with ethanol in discharges to surface water. The DEQ Valley Regional Office has reissued a permit to Merck & Co. to discharge treated production and sanitary wastewater generated at a pharmaceutical manufacturing facility, non-contact cooling water, and storm water generated in the area around the facility (Aschenbach, 2007). Revisions were made to the previous effluent limits, in part, so that new effluent monitoring and limitations matched the requirements of the Federal Effluent Guidelines for the Pharmaceutical Manufacturing Category. Although Virginia does not have a Water Quality Standard for ethanol, Outfall 101 of the Merck & Co. permit follows the EPA Guideline of 10 mg/L for a daily maximum limit (DML) and 4.1 mg/L for a monthly average limit (MAL) in terms of ethanol concentration, or 45 kg/d for a DML and 19 kg/d for an MAL in terms of ethanol loading. At the time of this writing, the Discharge Monitoring Report (DMR) analytical results for ethanol monitoring required to be performed once every six months are not yet due. The surface water that receives the discharge from the facility is designated as a Tier 1 water body which means that

the existing uses of the water body and water quality to protect such uses must be maintained in accordance with the State Water Control Board's antidegradation policy.

Ethanol does not bioaccumulate or bioconcentrate in the tissue of living organisms due to ethanol's chemical properties and to the ability of most organisms to metabolize ethanol (Iott, 2001). Human health risks from exposure to ethanol appear to be minimal, especially when compared with the risks posed by other gasoline constituents. Likewise, aquatic toxicity levels for ethanol are quite high. Ethanol also appears to degrade rapidly in both surface and subsurface environments. Based upon these factors, the DEQ does not believe that effluent limits for ethanol are needed for discharge of waters associated with petroleum products containing up to 10% ethanol.

Ethanol concentrations in discharges of petroleum products containing greater than 10% ethanol may pose risks to aquatic organisms. For discharge of petroleum products containing greater than 10% ethanol into surface water bodies not designated as a PWS, a maximum discharge limit of 4.1 mg/L is proposed. This same limit also is proposed for saltwater receiving bodies.

#### 6.1.10 pH

The pH limits in this general permit are based on the Virginia Water Quality Standards and range from a low of six (6.0) standard units to nine (9.0) standard units.

# 6.2 Basis for Effluent Limitations - Discharges of Petroleum Products other than Gasoline

The EPA model permit for UST remediation sites only addresses gasoline contaminated sites. This general permit is also designed to be used at sites which are contaminated by petroleum products other than gasoline (non-gasoline motor fuels, jet fuels, distillate fuel oils, residual fuel oils, lubricants, petroleum solvents and used oils). In addition to containing small amounts of the volatile organic compounds such as benzene, these products contain more of the polynuclear aromatic hydrocarbons (PAHs) than are found in gasoline. PAHs are less soluble in water than the volatile compounds and they are less amenable to air stripping. It is possible that a treatment system that is capable of removing the volatile compounds like benzene to acceptable levels may not effectively remove the PAHs. Based upon the types and relative proportions of the constituents present in the non-gasoline petroleum products, benzene and the BTEX constituents are not good indicator parameters to use in evaluating the quality of effluents from sites contaminated with this category of petroleum.

# 6.2.1 Naphthalene

The effluent limitation for naphthalene proposed in this general permit is a water quality-based limit. It is to be applied at sites where contamination is from petroleum products other than gasoline. Naphthalene is a component of gasoline and non-gasoline petroleum products, but its relative concentration is higher in products such as diesel and kerosene than in gasoline (Thomas & Delfino, 1991). It is less soluble in water than benzene (solubility 30 mg/l vs. 1780 mg/l) and is less amenable to air stripping (Henry's Law Constant  $4.83 \times 10^{-4}$  vs.  $5.55 \times 10^{-3}$  @  $25^{\circ}$ C). These characteristics make the treatability of naphthalene more similar to that of the heavier PAH components than the BTEX compounds.

PAHs in general are relatively insoluble in water. For instance, the solubilities of the typical petroleum PAHs anthracene, phenanthrene and fluorene are 1.29 mg/l, 0.8 mg/l and 1.9 mg/l, respectively. These compounds are more likely to be found in free product or adsorbed onto soils at a petroleum contaminated site rather than dissolved in groundwater. As a moderately soluble compound, naphthalene is more likely to dissolve in groundwater and migrate from the source of contamination. Therefore, it occupies an intermediate position between the volatile BTEX compounds and the less soluble PAHs. By selecting naphthalene as the indicator parameter for this category of contaminated sites, the general permit relies on the assumption that if naphthalene has been removed to acceptable levels, then the heavier PAHs associated with the contamination should have either remained in the soils at the source or been reduced to an acceptable level with the treatment for naphthalene.

The limited data available in the EPA Treatability Database indicate that treatment with granular activated

carbon (GAC) filtration is more effective in removing naphthalene and other PAHs than is air stripping. Although this general permit does not mandate a treatment technology, the low solubility of PAHs makes them amenable to treatment by GAC filtration of the contaminated groundwater.

The EPA criteria document for naphthalene (EPA 440/5-80-059) lists a chronic effect concentration of 620  $\mu$ g/l for fathead minnows, but it states that effects would occur at lower concentrations if more sensitive freshwater organisms were tested. According to the ECOTOX DATABASE, naphthalene at a concentration of 1,000  $\mu$ g/l was lethal to 50% of the water fleas (*Daphnia pulex*) tested (Truco et al., 1983). DeGaere and associates (1982) tested the effects of naphthalene on Rainbow Trout and reported an LC50 concentration of 1600  $\mu$ g/l. Based upon these more recent studies, it is recommended that the effluent limit for naphthalene in freshwater be set at 10  $\mu$ g/l.

The lowest observed LC50 value in the EPA criteria document for naphthalene (EPA, 1980e) reportedly was 2,350  $\mu$ g/l, in a test with grass shrimp. Korn and associates (1979) tested the effects of naphthalene on humpy shrimp (*Pandalus goniurus*) and found that a naphthalene concentration of 1020  $\mu$ g/l was lethal to 50% of the shrimp tested. Pink salmon (*Onchrhynchus gorbuscha*) were exposed to naphthalene and Rice and Thomas (1989) found that a concentration of 890  $\mu$ g/l was lethal to 50% of the fish tested. Dividing this LC50 by 100 results LC50 by 100 in the proposed saltwater effluent limit of 8.9  $\mu$ g/l.

There is no Virginia human health water quality standard for naphthalene. Equation (5) below is used by DEQ staff to derive human health based water quality standards for discharges of non-carcinogens to public water supplies. The human health derived value is much greater than the aquatic toxicity value of  $10~\mu g/l$ . It is recommended that freshwater aquatic toxicity value of  $10~\mu g/l$  be used for the naphthalene effluent limit in public water supplies.

Note: The reference dose is from the EPA IRIS database (EPA, 2007c) and the bioaccumulation factor is from EPA (2002).

# 6.2.2 Benzene and MTBE (discharges to a PWS only)

Benzene and MTBE are not found in high concentrations in petroleum products other than gasoline. MTBE is a gasoline additive and not intentionally placed in petroleum products other than gasoline. Benzene has a relatively low boiling point and most of the benzene in crude oil feedstocks will remain with the gasoline fraction hydrocarbons during the petroleum refining process.

After refining, petroleum products are transported via a common transportation network (pipelines, tanker trucks) and there is some unintentional mixing of products that occurs. While middle distillates (kerosene, diesel, #2 fuel oil) contain only very small amounts of benzene and MTBE is not intentionally placed in them, DEQ staff have found that MTBE and benzene are the most commonly found petroleum constituents in drinking water supplies contaminated by middle distillates. Due the presence of these constituents in water contaminated by petroleum products other than gasoline, it is recommended that all discharges of

petroleum-contaminated wastewater to public water supplies contain effluent limits for benzene and MTBE. Limits proposed for these constituents are 12 µg/l for benzene and 15 µg/l for MTBE.

# 6.2.3 Total Petroleum Hydrocarbons (TPH)

The general permit proposes a technology-based limit of 15 mg/l for the parameter Total Petroleum Hydrocarbons (TPH). This limit is applicable for discharges where the contamination is from petroleum products other than gasoline. It is based on the ability of simple oil/water separator technology to recover free product from water. Wastewater that is discharged without a visible sheen is generally expected to meet this effluent limitation. Monitoring data generated during a previous term of general permit VAG83 indicates that effluents are generally below this level. DEQ has utilized an effluent limitation of 15 mg/l oil & grease for many years in individual permits for potential sources of petroleum hydrocarbons. Recently, the DEQ determined that the oil & grease analytical method is better suited for detection of animal and vegetable fats rather than petroleum. Therefore, the parameter TPH is being used in the general permit rather than oil & grease.

The term "used oils" is used in the general permit to refer to those petroleum products that have served their useful purpose and have been collected for recycling or disposal. Tanks that store used oils are found at industrial sites and at automotive service stations. These tanks have the potential to leak into surrounding soils and contaminate groundwater. The materials in used oil storage tanks can be a mixture of motor oils and other petroleum products, as well as solvents or other organic chemicals. Used oils also may contain dissolved metals derived from the machinery from which the oil was recovered. These mixtures pose potential environmental impacts that may not be adequately addressed by the pollutant parameters established to control discharges from the sites contaminated by products other than gasoline. Therefore, the general permit proposes to require that when the contamination is from used oils, additional monitoring shall be conducted to scan the wastewater for a wide range of organic compounds and metals. This information will be evaluated and a decision on the need for additional limits on discharges of this type will be made prior to the expiration date of the general permit. In no case will the general permit allow a discharge of wastewaters if the contamination is from used oils that are classified as hazardous materials according to the Virginia Hazardous Waste Regulation, 9 VAC 20-60.

# 6.3 Discharges from Hydrostatic Testing of Tanks and Pipelines

When this permit was reissued in 1998, hydrostatic test waters from petroleum facilities were included so that a VPDES permit could properly govern them. The permit regulation was further expanded in 2003 to include coverage of discharges from hydrostatic testing of natural gas pipelines.

Natural gas, like other petroleum products, is not constant in its composition or the relative proportions of individual constituents within that product. According to Technocarb (2002), methane typically makes up approximately 95 percent of natural gas by volume. Ethane and propane generally make up approximately two and one percent of the gas, respectively. Other constituents that typically make up the remaining two percent of the mixture include butane, carbon dioxide, and nitrogen. There is no aquatic or human toxicity data for these compounds.

Discharges from hydrostatic testing of pipelines are generally one-time occurrences of less than 48 hours. Such frequencies and durations preclude the necessity for application of toxic parameters except for total residual chlorine (TRC). TRC is potentially present in high concentrations when treated potable water is used as the source water for testing. Discussion of the recommended effluent limits for discharges of hydrostatic test water from natural gas pipelines is presented below. In addition to the effluent limits, the following requirements will also apply to hydrostatic discharges from natural gas pipelines:

- 1. The equipment being tested shall be substantially free of debris, raw material, product, or other residual materials.
- 2. The discharge flow shall be controlled in such a manner that prevents flooding, erosion, or excessive sediment influx into the receiving water body.

# **6.3.1** Total Petroleum Hydrocarbons (TPH)

The limit for TPH is based on the ability of simple oil-water separator technology to recover petroleum from water. Wastewater that is discharged without a visible sheen is generally expected to meet this effluent limitation. DEQ has used this limitation for many individual permits for many years and monitoring data has demonstrated that it is readily achievable. Mass limits are not applicable to this type of pollutant and discharge and are not required.

#### 6.3.2 Total Organic Carbon (TOC)

Total organic carbon (TOC) is monitored to assure that the effluent is not contaminated with non-petroleum organic substances. Staff members generally believe that TOC concentrations in this type of discharge are low. However, should sampling data indicate high levels of TOC, the permit may be modified at a later time to include such a limit.

# 6.3.3 Total Suspended Solids (TSS)

Total suspended solids (TSS) is monitored to assure that the effluent is not contaminated with excessive amounts of solids that might be flushed out of pipes along with the test waters. If significant concentrations of suspended solids are detected, the permit may be modified at a later time to include a limit.

#### 6.3.4 Total Residual Chlorine (TRC)

Total residual chlorine (TRC) is necessary for those hydrostatic tests that use chlorinated potable drinking water as the source water for testing. The limit is based on the chronic aquatic life criterion in Virginia's water quality standards.

#### 6.3.5 pH

The pH limits in this general permit are based on the Virginia Water Quality Standards and range from six (6.0) standard units to nine (9.0) standard units.

# 6.4 Discharges of Water Contaminated by Chlorinated Hydrocarbon Solvents

Many different chlorinated hydrocarbons are, or have been, used as solvents. Dealing with these materials when they have been released into the environment is further complicated by the fact that they often break down into other chlorinated hydrocarbon compounds; many of which also are solvents. Therefore, although only one type of chlorinated hydrocarbon may have been released at a site, subsequent cleanup efforts may have to deal with multiple chlorinated hydrocarbons. Figures 1 and 2 show the degradation products that are or can be created by the breakdown of 1,1,1 trichloroethane, tetrachloroethane, and carbon tetrachloride.

Effluent limits recommended for chlorinated hydrocarbon solvent constituents were based upon both the toxicity of the material as well as treatment technology. Some of the toxicity-based limits that were considered include promulgated water quality standards, drinking water maximum contaminant levels (MCLS), aquatic toxicity data from the EPA ECOTOX database, and tap water risk—based concentrations from EPA Region III. Staff also considered effluent limits that had been placed in individual VPDES permits.

Staff recommended one set of effluent limits for these chlorinated hydrocarbon solvents and set the limits to protect both aquatic life and human health. The effluent limits were based upon the assumption of a discharge into a PWS and the limits had to meet criteria for public water supplies. Table 1 summarizes the pertinent regulatory values that exist for chlorinated hydrocarbon solvent compounds and the effluent limits that have been proposed for these constituents.

Figure 1. Reductive Dehalogenation of 1,1,1 TCA and Tetrachloroethylene (from Dragun 1988)

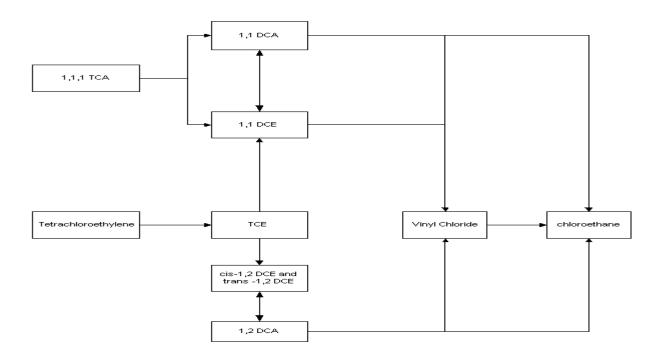
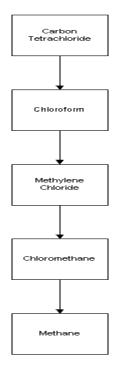


Figure 2. Reductive Dechlorination of Carbon Tetrachloride (from RTDF Bioremediation Consortium 1988)



Fact Sheet: General VPDES Permit for Discharges from Petroleum Contaminated

Sites, Groundwater Remediation, and Hydrostatic Tests, VAG83

Name	CAS Number	Effluent limits from individual permits (µg/l)	Drinking Water MCL (μg/l)	WQS, HH for PWS <sup>1</sup> (μg/l)	WQS, HH for Other Waters <sup>2</sup> (μg/l)	Toxicity FW³ (μg/l)	Toxicity SW³ (μg/l)	EPA Reg. III Tap Water RBC <sup>4</sup> (μg/I)	Recommended Effluent Limit (µg/l)
Chloroform	67663	100 (3 permits)		350	29000	290	815		100
1,1 Dichloroethane	75343	4 (one permit), 5 (2 permits)						900	4
1,2 Dichloroethane	107062	5 (3 permits)	5	3.8	990	1160	1130		3.8
1,1 Dichloroethylene <sup>A</sup>	75354	7 (4 permits)	7	310	17000	740	2240		7
cis-1,2 Dichloroethylene	159592	70 (3 permits)	70						70
trans-1,2 Dichloroethylene	156605	100 (4 permits)	100	700	140000	2200			100
Methylene Chloride <sup>A</sup>	75092	5 (2 permits)	5	47	16000	1930	770		5
Tetrachloroethylene <sup>A</sup>	127184	5 (4 permits) and 79 (1 permit)	5	8	89	18	13		5
1,1,1 Trichloroethane	71556	200 (4 permits)	200			112	3120		112
1,1,2 Trichloroethane	79005	5	5	6	420	180	270		5
Trichloroethylene	79016	5 (3 permits)	5	27	810	19	140		5
Vinyl Chloride	75014	2 (3 permits)	2	0.23	61				2
Carbon Tetrachloride	56235	5	5	2.5	44	20	500		2.5
1,2 Dichlorobenzene	95501	600		2700	17000	15.8	19.7		15.8
Chlorobenzene	108907	NL	100	680	21000	3.4	89		3
Trichlorofluoromethane	75694	5						1300	5
Chloroethane <sup>A</sup>	75003	5						3.6	3

<sup>&</sup>lt;sup>1</sup> The values in this column are human health criteria for public water supplies from the Virginia Water Quality Standards (9 VAC 25-260).

<sup>&</sup>lt;sup>2</sup> The values in this column are human health criteria for surface waters that are not public water supplies. These numbers are from the Virginia Water Quality Standards (9 VAC 25-260).

Aquatic toxicity values were derived from the EPA ECOTOX database.

These are tap water risk-based concentrations from the EPA Region III Risk-Based Concentration Table. These values are provided only for constituents for which regulatory concentrations do not exist.

Synonyms: dichloromethane = methylene chloride; ethyl chloride = chloroethane; 1,1 dichloroethene = 1,1 dichloroethylene; perchloroethylene = tetrachloroethylene

# 6.4.1 Chloroform

According to Howard (1990), chloroform is used as an industrial solvent, extractant, and chemical intermediate. Chloroform also may be created by the reductive dehalogenation of carbon tetrachloride that has been released into the environment (RRDF Bioremediation Group, 1988). The human-health Water Quality Standards for chloroform are 350 μg/l for public water supplies and 29,000 μg/l for other surface waters. The DEQ Northern Regional Office has issued three individual permits having an effluent limit for chloroform and Northern Regional Staff used a technology-based limit of 100 μg/l for all three permits. LeBlanc (1980) found that chloroform, at a concentration of 29000 μg/l, killed fifty percent of the water fleas (*Daphnia magna*) tested. Bentley and associates (1979) found that chloroform killed fifty percent of the pink shrimp (*Penaeus douranum*) tested when the chloroform concentration was 81500 μg/l. Applying the safety factor of 100 to these LC50 values resulted in chronic toxicity levels for freshwater and saltwater organisms of 290 and 815 μg/l respectively. The technology-based limit of 100 μg/l that was used for the individual VPDES permits in the Northern Region is the most conservative and protective concentration and is recommended as the effluent limit for chloroform.

#### 6.4.2 1,1 Dichloroethane

1,1 Dichloroethane (1,1 DCA) predominantly is used to make other chemicals (Howard, 1990, and ATSDR, 1999a). This constituent also is used to dissolve substances such as paint and varnish, and as a degreasing agent (ATSDR, 1999a). 1,1 DCA may be created by the breakdown of 1,1,1 trichloroethane that has been released into the environment (Dragun, 1988). There is very limited aquatic toxicity information for 1,1 dichloroethane. There are no promulgated drinking water standards for this constituent nor is there a drinking water MCL. The EPA ECOTOX database contains no information for this constituent. The EPA Region III risk-based concentration for this constituent in tap water is 900  $\mu$ g/l. The DEQ Northern Regional Office has placed an effluent limit of 4  $\mu$ g/l for this constituent in one individual VPDES permit and 5  $\mu$ g/l in two permits. The TAC recommends an effluent limit of 4  $\mu$ g/l for 1,1 dichloroethane.

# 6.4.3 1,2 Dichloroethane

According to ATSDR (2001a), 1,2 dichloroethane (1,2 DCA) is used in the production of vinyl chloride which, in turn, is used to make a variety of plastic and vinyl products. 1,2 DCA also is used as a solvent and as a lead scavenger in leaded gasoline. This constituent may be created in the environment by reducing the carbon-carbon double bonds in the cis and trans 1,2 dichloroethylene isomers (Dragun, 1988). The Northern Regional Office has placed an effluent limit of 5  $\mu$ g/l for 1,2 dichloroethane (1,2 DCA) in 3 individual VPDES permits. The Federal drinking water MCL for 1,2 DCA is 5  $\mu$ g/l. Virginia's human-health based water quality standards for this constituent are 3.8  $\mu$ g/l and 990  $\mu$ g/l for public water supplies and for other surface waters, respectively. According to the ECOTOX database, the lowest saltwater LC50 concentration for 1,2 DCA is 113000  $\mu$ g/l (EPA, 1978). The lowest freshwater LC50 concentration reported for 1,2 DCA is 116000  $\mu$ g/l (Walbridge, 1983). Applying the safety factor of 100 to these LC50 values results in concentrations of 1160  $\mu$ g/l and 1130  $\mu$ g/l for freshwater and saltwater, respectively. The water quality criteria of 3.8  $\mu$ g/l for public water supplies is more protective than the drinking water MCL and the aquatic toxicity-based values and is recommended as the effluent limit.

#### 6.4.4 1,1 Dichloroethylene

1,1 Dichloroethylene (1,1 DCE) is used in the manufacture of plastic wrap, adhesives, and synthetic fiber (Howard, 1989). This constituent also is formed during the anaerobic biodegradation of trichloroethylene (TCE) and the hydrolysis of 1,1,1 trichloroethane (1,1,1 TCA, Howard, 1989, and Dragun, 1988). The human health Water Quality Standards for 1,1 DCE are 310 µg/l for public water supplies and 17000 µg/l for other surface waters. The Federal drinking water MCL for 1,1 DCE is 7 µg/l. Dill and associates (1980) found that 1,1 DCE at a concentration of 11600 µg/l killed half of the water fleas (*Daphnia magna*) tested. The lowest reported LC50 concentration for saltwater organisms was 224000 µg/l (EPA 1978). The DEQ Northern Regional Office has an effluent limit of 7 µg/l for 1,1 DCE in four individual VPDES permits. This effluent limit is the same as the Federal MCL and is recommended as the effluent limit for this general

permit.

### 6.4.5 cis-1,2 Dichloroethylene

The cis-1,2 dichloroethylene (cis 1,2 DCE) isomer is not a priority pollutant. Much of the cis-1,2 DCE that is found in the environment comes from reductive dehalogenation of trichloroethylene (Howard, 1990). The Federal MCL for cis-1,2 DCE is 70  $\mu$ g/l. The DEQ Northern Regional Office has three individual VPDES permits with effluent limits for this constituent and all of them have an effluent limit of 70  $\mu$ g/l. The TAC recommends an effluent limit of 70  $\mu$ g/l for cis-1,2 DCE.

#### 6.4.6 trans 1,2 Dichloroethylene

Trans 1,2 dichloroethylene (trans-1,2 DCE) is a priority pollutant and the preferred isomer of DCE in most industrial applications (HSDB, 1995). This constituent is used as a solvent and extractant and also is used in manufacturing perfumes, lacquers, and thermoplastics (Howard, 1990). Trans-1,2 DCE also can be created by the reductive dehalogenation of trichloroethylene (Dragun, 1988). The Federal drinking water MCL for trans-1,2 DCE is 100  $\mu$ g/l. Northern Regional Office staff also used an effluent limit of 100  $\mu$ g/l for trans-1,2 DCE in four individual VPDES permits issued by that office. Human health-based water quality standards for this constituent are 700  $\mu$ g/l for public water supplies and 140,000  $\mu$ g/l for other surface waters. LeBlanc (1980) found that a concentration of 220,000  $\mu$ g/l trans-1,2 DCE in water was lethal to 50 percent of the water fleas (*Daphnia magna*) tested. The TAC recommends that the effluent limit for trans-1,2 DCE be set at 100  $\mu$ g/l.

#### 6.4.7 Methylene Chloride

Methylene chloride is used as a solvent and paint remover, may be found in certain aerosols and pesticides, and is used to manufacture photographic film (Howard, 1990, and ATSDR, 2001b). According to the RTDF Bioremediation Consortium (1998), methylene chloride also may be derived from the anaerobic degradation of chloroform. The lowest freshwater LC50 concentration reported for methylene chloride is 193000  $\mu$ g/l for fathead minnows (*Pimephales promelas*, Alexander, 1978). Burton and Fisher (1990) found that methylene chloride, at a concentration of 97000  $\mu$ g/l, was lethal to 50 percent of the mummichogs (*Fundulus heteroclitus*) tested. The Federal drinking water MCL for methylene chloride is 5  $\mu$ g/l and this is also the effluent limit that the Northern Regional Office staff used in the two permits that have limits for this constituent. The Water Quality Standards for methylene chloride are 47  $\mu$ g/l and 16000  $\mu$ g/l for public water supplies and other surface waters, respectively. The TAC recommends an effluent limit of 5  $\mu$ g/l for methylene chloride.

#### 6.4.8 Tetrachloroethylene

Tetrachloroethylene, also known as perchloroethylene, is used widely for dry cleaning fabrics and as a metal degreasing agent (Howard, 1990, and ATSDR, 1997). According to Yoshioka and others (1986), tetrachloroethylene at a concentration of 1800  $\mu$ g/l was lethal to 50 percent of the water fleas (*Moina macrocopa*) tested. The lowest saltwater LC50 value reported for tetrachloroethylene is 1300  $\mu$ g/l for daggerblade grass shrimp (*Palaemonetes pugio*, Horne et al., 1983). Applying the safety factor of 100 to these LC50 values results in limits of 18  $\mu$ g/l and 13  $\mu$ g/l, respectively. The human health-based water quality standards for tetrachloroethylene are 8  $\mu$ g/l for public water supplies and 47  $\mu$ g/l for other surface waters. The Federal drinking water MCL for tetrachloroethylene is 5  $\mu$ g/l. Five individual VPDES permits in the Northern Regional Office have effluent limits for tetrachloroethylene. Four of these permits have an effluent limit of 5  $\mu$ g/l and one of the permits has an effluent limit of 79  $\mu$ g/l. The TAC recommends an effluent limit of 5  $\mu$ g/l for tetrachloroethylene.

# 6.4.9 1,1,1 Trichloroethane

1,1,1 Trichloroethane (1,1,1 TCA) formerly was used as a solvent to dissolve glues and paints, a degreasing agent for metal parts, and is an ingredient of household products such as glues, spot removers, and aerosol sprays (ATSDR, 2006a, and Howard, 1990). According to ATSDR 2006a, TCA was not supposed to be

manufactured for domestic use in the United States after January 1, 2002, due to its effects on the ozone layer. The Federal drinking water MCL for 1,1,1 Trichloroethane (1,1,1 TCA) is 200 μg/l. Four individual VPDES permits in the Northern Regional Office have effluent limits for 1,1,1 TCA and the effluent limit in each permit is 200 μg/l. Virginia does not have promulgated water quality standards for 1,1,1 TCA. The lowest freshwater LC50 value for 1,1,1 TCA that is reported in the ECOTOX database is 11200 μg/l for water fleas (*Daphnia magna*, Cowgill, 1987). EPA (1978) found that 1,1,1 TCA at a concentration of 312000 was lethal to 50 percent of the opossum shrimp (*Americamysis bahia*) tested. If the customary safety factor of 100 is applied to these LC50 values, results in concentrations of 112 μg/l and 3120, respectively that are expected to be protective of aquatic and marine life. The most conservative or protective concentration for 1,1,1 TCA is the value that was derived from toxicity of this constituent to water fleas. The TAC recommends an effluent limit of 112 μg/l for 1,1,1 TCA.

#### 6.4.10 1,1,2 Trichloroethane

1,1,2 TCA is a solvent and an intermediate in the production of 1,1 DCA (ATSDR 199b). Only one individual permit in the Northern Regional Office has an effluent limit for 1,1,2 TCA and the limit in that permit is 5  $\mu$ g/l. The Federal drinking water MCL for 1,1,2 TCA also is 5  $\mu$ g/l. The Virginia Water Quality Standards for 1,1,2 TCA are 6  $\mu$ g/l for public water supplies and 420  $\mu$ g/l for other surface waters. LeBlanc (1980) found that 1,1,2 TCA, at a concentration of 18,000  $\mu$ g/l, was lethal to 50 percent of the water fleas (*Daphnia magna*) tested. The lowest LC50 value reported for this constituent for saltwater organisms is 27,000  $\mu$ g/l (Adema and Vink, 1981). Applying the safety factor of 100 to these LC50 values results in concentration of 18  $\mu$ g/l and 27  $\mu$ g/l, respectively. The TAC recommends an effluent limit of 5  $\mu$ g/l for 1,1,2 TCA.

# 6.4.11 Trichloroethylene

Trichloroethylene (TCE) is a solvent commonly used to remove grease from metal parts (Howard, 1990, and ATSDR, 2003). TCE also is an ingredient in certain adhesives, paint removers, typewriter correction fluids, and spot removers (ATSDR, 2003). TCE can be formed by the breakdown of tetrachloroethylene that has been released into the environment. The Federal drinking water MCL for TCE is 5  $\mu$ g/l and this is the same effluent limit that the Northern Regional Office staff used for all three VPDES permits that contained limits for TCE. The promulgated water quality standard for public water supplies is 27  $\mu$ g/l and the water quality standard for all other surface water is 810  $\mu$ g/l. The lowest freshwater LC50 value reported to TCE is 1900  $\mu$ g/l (Yoshioka, 1986). Ward and associates (1986) found that TCE at a concentration of 14000  $\mu$ g/l was lethal to 50 percent of the opossum shrimp (*Americamysis bahia*) tested. Applying the safety factor of 100 to these LC50 values results in concentrations of 19  $\mu$ g/l and 140  $\mu$ g/l. The TAC recommends an effluent limit of 5  $\mu$ g/l for TCE.

# 6.4.12 Vinyl Chloride

Most vinyl chloride is used to manufacture polyvinyl chloride (PVC, Howard, 1989, and ATSDR, 2006b). This constituent generally is not used as a solvent, but it is commonly found in the environment due the breakdown of other chlorinated hydrocarbon solvents (Dragun, 1988, and ATSDR, 2006b). The Federal drinking water MCL for vinyl chloride is 2  $\mu$ g/l and this is the effluent limit that the DEQ Northern Regional Office staff have used for all three of their individual VPDES permits having a limit for this constituent. The Water Quality Standard for public water supplies is .23  $\mu$ g/l and the water quality standard for other surface waters is 61  $\mu$ g/l. The TAC recommends an effluent limit of 2  $\mu$ g/l for vinyl chloride. This limit is the same as the drinking water MCL and, as a promulgated MCL, is both protective and achievable. Current analytical methods typically cannot quantify vinyl chloride or other volatile organic compounds at concentrations of less than 1  $\mu$ g/l. MCLs are set at limits that are believed protective of human health and can be reached by current treatment technologies. Members of the TAC are not confident that an effluent limit of 2  $\mu$ g/l for vinyl chloride may be achieved by current treatment technologies.

#### 6.4.13 Carbon Tetrachloride

According to Howard (1990) large quantities of carbon tetrachloride are used for the chemical synthesis of fluorocarbon refrigerants and propellants. Carbon tetrachloride also is used as a degreaser, a cleaning fluid, and a grain fumigant pesticide (Howard, 1990, and ATSDR, 2005). The Water Quality Standards for carbon tetrachloride are 2.5  $\mu$ g/l for public water supplies and 44  $\mu$ g/l for other surface waters. The Federal drinking water MCL for carbon tetrachloride is 5  $\mu$ g/l. DEQ staff in the Northern Regional Office have issued one individual VPDES permit having an effluent limit for carbon tetrachloride and that limit was 5  $\mu$ g/l. Yoshioka and associates (1986)found that carbon tetrachloride at a concentration of 2000  $\mu$ g/l was lethal to 50 percent of the Medaka, high-eyes (*Oryzias latipes*) tested. The lowest saltwater LC50 value listed in the ECOTOX database was 50,000  $\mu$ g/l for sole order (*Pleuronectiformes*, Pearson and McConnell, 1975). The TAC recommends an effluent limit of 2.5  $\mu$ g/l for carbon tetrachloride.

#### 6.4.14 1,2 Dichlorobenzene

According to the National Toxicology Program (NTP), U.S. Department of Health and Human Services (1985), the major use of 1,2 dichlorobenzene is as an intermediate in the synthesis of other organic compounds including the herbicides propanil, diuron, and neburon. This constituent also is used as an engine cleaner and de-inking solvent, a degreasing agent, a heat exchange medium, and a fumigant pesticide (NTP 1985). The water quality standard for 1,2 dichlorobenzene in public water supplies is 2700 μg/l and the water quality standard for other surface waters is 17,000 μg/l. There is no promulgated Federal drinking water MCL for this constituent. Staff in the Northern Regional Office issued one individual VPDES permit having an effluent limit for 1,2 dichlorobenzene and the limit in that permit was 600 μg/l. EPA (1978) reported that 1,2 dichlorobenzene at a concentration of 1970 μg/l killed 50 percent of the opossum shrimp (*Americamysis bahia*) tested. The lowest freshwater LC50 value reported in the ECOTOX database for this constituent was 1580 μg/l for rainbow trout (*Oncorhynchus mykiss*, Call and Associates, 1983). Applying the customary safety factor of 100 to the LC50 value for rainbow trout results in a concentration of 15.8 μg/l. The TAC recommends an effluent limit of 15.8 μg/l for 1,2 dichlorobenzene.

#### 6.4.15 Chlorobenzene

Chlorobenzene production has declined by over half since its peak of use in 1960 (ATSDR, 1998). Presently, chlorobenzene is used as a solvent for certain pesticides, a degreasing agent for automobile parts, and a chemical intermediate to make other chemicals (ATSDR, 1998). The Federal drinking water MCL for chlorobenzene is 100 µg/l. The water quality standards for this constituent are 680 µg/l for public water supplies and 21,000 µg/l for other surface waters. Birge and others (1979) reported that a concentration of 340 µg/l was lethal to 50 percent of the largemouth bass (*Micropterus salmoides*) they tested. The lowest saltwater LC50 value reported in the ECOTOX database for this constituent is 8900 µg/l for sheepshead minnows (*Cyprinodon variegates*, Heitmuller and others, 1981). Applying the customary safety factor of 100 to these LC50 values results in concentrations of 3.4 µg/l and 89 µg/l, respectively. The TAC recommends an effluent of 3.4 µg/l for chlorobenzene.

#### 6.4.16 Trichlorofluoromethane

Trichlorofluoromethane, also known as Freon 11, was used as a propellant for aerosol sprays until its use for this application was banned in the United States on December 15, 1978 (Howard, 1990). Trichlorofluoromethane also is used as a refrigerant, foaming agent for polyurethane foams, solvent and degreaser, and fire extinguishing agent (Howard, 1990). Limited information exists for trichlorofluoromethane. There is no MCL for this constituent, no promulgated water quality standards, and no aquatic toxicity data that has been summarized in the ECOTOX database. The DEQ Northern Regional Office staff have written one individual permit having an effluent limit for this constituent and that effluent limit is  $5 \mu g/l$ . EPA Region III has listed a risk-based value for trichlorofluoromethane in tap water and that concentration is  $1300 \mu g/l$ . The TAC recommends an effluent limit of  $5 \mu g/l$  for trichlorofluoromethane.

#### 6.4.17 Chloroethane

According to ATSDR (1999c), chloroethane is used in the production of cellulose dyes, medicinal drugs,

and other commercial products. This constituent also is used as a solvent and refrigerant. Chloroethane is used to numb the skin prior to ear piercing and skin biopsies and also as a treatment for sports injuries (ATSDR, 1999c). Chloroethane has been shown to form as a degradation byproduct of other chlorinated hydrocarbon solvents (Howard, 1990, and Dragun, 1988). Like trichlorofluoromethane, little aquatic toxicity information exists for chloroethane. The DEQ Northern Regional Office staff have written one individual permit having an effluent limit for this constituent and that effluent limit is  $5 \mu g/l$ . EPA Region III has listed a risk-based value for chloroethane in tap water and that concentration is  $3.6 \mu g/l$ . The TAC recommends an effluent limit of  $3.6 \mu g/l$  for chloroethane.

# 7.0 Administration of this General Permit Regulation

The general permit will have a fixed term of five (5) years effective upon Board approval. Every authorization to discharge under this general permit will expire at the same time and all authorizations to discharge will be renewed on the same date. Discharges will be covered under the general permit upon approval of the Registration Statement and delivery of a copy of the general permit to the applicant.

This general permit does not apply to any new or increased discharge that will result in significant effects to the receiving waters. That determination is made in accordance with the State Water Control Board's Antidegradation Policy contained in the Virginia Water Quality Standards, 9 VAC 25-260. Antibacksliding will also be considered prior to granting coverage under this general permit to operations currently discharging under another VPDES permit.

If an applicant for a discharge appears to qualify for this general permit, the applicant will be required to submit a general permit Registration Statement. The Board will review the Registration Statements received and either send a copy of the general permit to those that qualify, or send a copy of the application for an individual permit to those that do not qualify.

#### REFERENCES

Adema, D.M.M. and G.J. Vink. 1981. "A Comparative Study of the Toxicity of 1,1,2 Trichloroethane, Dieldrin, Pentachlorophenol, and 3,4 Dichloroaniline for Marine and Fresh Water Organisms." Chemisphere. 10(6): 533-554.

Agency for Toxic Substances and Disease Registry (ATSDR). 1997. ToxFAQs for Tetrachloroethylene.

Agency for Toxic Substances and Disease Registry (ATSDR). 1998. ToxFAOs for Chlorobenzene.

Agency for Toxic Substances and Disease Registry (ATSDR). 1999a. ToxFAQs for 1,1 Dichloroethane.

Agency for Toxic Substances and Disease Registry (ATSDR). 1999b. ToxFAQs for 1,1,2 Trichloroethane.

Agency for Toxic Substances and Disease Registry (ATSDR). 1999c. ToxFAQs for Chloroethane.

Agency for Toxic Substances and Disease Registry (ATSDR). 2001a. ToxFAQs for 1,2 Dichloroethane.

Agency for Toxic Substances and Disease Registry (ATSDR). 2001b. ToxFAQs for Methylene Chloride.

Agency for Toxic Substances and Disease Registry (ATSDR). 2003. ToxFAQs for Trichloroethylene.

Agency for Toxic Substances and Disease Registry (ATSDR). 2005. ToxFAQs for Carbon Tetrachloride.

Agency for Toxic Substances and Disease Registry (ATSDR). 2006a. ToxFAQs for 1,1,1 Trichloroethane.

Agency for Toxic Substances and Disease Registry (ATSDR). 2006b. ToxFAQs for Vinyl Chloride.

Alexander, H.C., W.M. McCarty, and E.A. Bartlett. 1978. "Toxicity of Perchloroethylene, Trichloroethylene, 1,1,1 Trichloroethane, and Methylene Chloride to Fathead Minnows." Bull. Environ. Contam. Toxicol. 20(3): 344-352.

Aschenbach, E.T., 2007, Personal communication (email). Virginia Department of Environmental Quality-Valley Regional Office, Harrisonburg, Virginia.

- BenKinney, M.T., J.F. Barbieri, J.S. Gross, and P.A. Naro. 1994. Aquatic toxicity of methyl-tertiary-butyl ether (MTBE) to aquatic organisms. Presented at the 15<sup>th</sup> Annual SETAC meeting, October 30, 1994. Denver. Co.
- Bentley, R.E., T. Heitmuller, B.H. Sleight III, and P.R. Parrish. 1979. Acute toxicity of Chloroform to Bluegill (*Lepomis macrochirus*), Rainbow Trout (*Salmo gairdneri*), and Pink Shrimp (*Penaeus dourarum*). U.S. EPA Criteria Branch, WA-6-99-1414-B, Washington, D.C.
- Birge, W.J., J.A. Black, and M.M. Bruser. 1979. Toxicity of Organic Chemicals to Embryo-Larval Stages of Fish. EPA-560/11-79-007, U.S. EPA, Washington, D.C.
- Boeri, R.L., J.P. Magazu and T.J. Ward. 1994. Acute toxicity of methyl tertiary butyl ether to the mysid, *Mysidopsis bahia*. Study No. 424-AR. T.R. Wilbury Laboratories, Inc., MA.
- Burton, D.T. and D.J. Fisher. 1990. "Acute Toxicity of Cadmium, Copper, Zinc, Ammonia, 3,3" dichlorobenzidine, 2,6 Dichloro-4-Nitroaniline, Methylene Chloride, and 2,4,6 Trichlorophenol to Juvenile Grass Shrimp and Killifish." Bull. Environ. Contam. Toxicol. 44(5): 776-783.
- Cowgill, U.M. 1987. "Critical Analysis of Factors Affecting the Sensitivity of Zooplankton and the Reproducibility of Toxicity Test Results." Water Res. 21(12): 1453-1462.
- Davis, J.T. and W.W. Hardcastle. 1959. "Biological Assay of Herbicides for Fish Toxicity." Weeds. 7: 397-404.
- DeGrave, G.M., R.G. Elder, D.C. Woods, and H.L. Bergman. 1982. "Effects of Naththalene and Benzene on Fathead Minnows and Rainbow Trout." Arch. Environ. Contam. Toxicol. 11(4): 487-490.
- Dragun, J. 1988. The Soil Chemistry of Hazardous Materials. Hazardous Materials Control Research Institute, Silver Spring, MD.
- EPA. 1978. In-Depth Studies on Health and Environmental Impacts of Selected Water Pollutants. U.S. EPA Contract No. 68-01-4646, Duluth, MN.
- EPA. 1980a. Ambient Water Quality Criteria for Benzene. EPA 440/5-80-018. USEPA Office of Water Regulations and Standards, Criteria and Standards Division, Washington, D.C.
- EPA. 1980b. Ambient Water Quality Criteria for Ethylbenzene. EPA 440/5-80-048. USEPA Office of Water Regulations and Standards, Criteria and Standards Division, Washington, D.C.
- EPA. 1980c. Ambient Water Quality Criteria for Toluene. EPA 440/5-80-075. USEPA Office of Water Regulations and Standards, Criteria and Standards Division, Washington, D.C.
- EPA. 1980d. Ambient Water Quality Criteria for Chlorinated Ethanes. EPA 440/5-80-029. USEPA Office of Water Regulations and Standards, Criteria and Standards Division, Washington, D.C.
- EPA. 1980e. Ambient Water Quality Criteria for Naphthalene. EPA 440/5-80-059. USEPA Office of Water Regulations and Standards, Criteria and Standards Division, Washington, D.C.
- EPA. 1999. Analytical Method Guidance for the Pharmaceutical Manufacturing Point Source Category. EPA Office of Water. EPA 821-B-99-003.
- EPA. 2000. "Methyl tertiary-butyl ether (MTBE); Advance notice of intent to initiate rulemaking under the Toxic Substances Act to eliminate or limit the use of MTBE as a fuel additive in gasoline." Federal Register 65(68):16093-16109.
- EPA. 2002. National Recommended Water Quality Criteria: 2002, Human Health Criteria Calculation Matrix. EPA-822-R-02-012.
- EPA Region III. 2006. Risk Based Concentration Table. <a href="http://www.epa.gov/reg3hwmd/risk/human/index.htm">http://www.epa.gov/reg3hwmd/risk/human/index.htm</a>

Fact Sheet: General VPDES Permit for Discharges from Petroleum Contaminated Sites, Groundwater Remediation, and Hydrostatic Tests, VAG83

- EPA. 2006. Fact Sheet: Ambient Water Quality Criteria Update for Methyl Tertiary-Butyl Ether (MTBE). EPA -822-F-06-002.
- EPA. 2007a. Webpage on Electronic Code of Federal Regulations (e-CFR). Title 40, Chapter I, Subchapter N, PART 414-ORGANIC CHEMICALS, PLASTICS, AND SYNTHETIC FIBERS.
- EPA. 2007b. Webpage on Electronic Code of Federal Regulations (e-CFR). Title 40, Chapter I, Subchapter N, PART 439, PHARMACEUTICAL MANUFACTURING POINT SOURCE CATEGORY.
- EPA. 2007c. Integrated Risk Information System (IRIS) Database. http://www.epa.gov/iris
- EXTOXNET. 1996. Extension Toxicology Network, Pesticide Information Profiles. Oregon State University
- Geiger, D.L., D.J. Call, and L.T. Brooke. 1988. Acute Toxicities of Organic Chemicals to Fathead Minnows (Pimephales promelas), Vol. 4. Center for Lake Superior Environmental Studies, University of Wisconsin, Superior.
- Fujiwara, Y. et al. 1984. "Biodegradation and bioconcentration of alkyl ethers." Yukayaken. 33:111-114. (original paper unavailable, available version translated by SCITRAN, Santa Barbara, CA.).
- Heitmuller, P.T., T.A. Hollister, and P.R. Parrish. 1981. "Acute Toxicity of 54 Industrial Chemicals to Sheepshead Minnows (Cyprinodon variegates)." Bull. Enviorn. Contam. Toxicol. 27(5): 596-604.
- Horne, J.D., M.A. Swirsky, T.A. Hollister, B.R. Oblad, and J.H. Kennedy. 1983. Aquatic Toxicity Studies of Five Priority Pollutants. Rep. No. 4398, Final Report, EPA Contract No. 68-01-6201, NUC Corp., Houston, TX.
- Howard, Philip H. (ed.) 1989. Handbook of Environmental Fate and Exposure Data for Organic Chemicals. Volume I Large Production and Priority Pollutants. Lewis Publishers, Boca Raton, FL.
- Howard, Philip H. (ed.) 1990. Handbook of Environmental Fate and Exposure Data for Organic Chemicals. Volume II Solvents. Lewis Publishers, Boca Raton, FL.
- HSDB. 1995. Hazardous Substances Databank (online database). National Laboratory of Medicine, National Toxicology Information Program. Bethesda, MD.
- Hwang, J., 2007, Personal communication, United States Environmental Protection Agency Region III, Philadelphia, Pennsylvania.
- Korn, S., D.A. Moles, and S.D. Rice. 1979. "Effects of Temperature on the Median Tolerance Limit of Pink Salmon and Shrimp Exposed to Toluene, Naphthalene, and Cook Inlet Crude Oil." Bul. Environ. Contam. Toxicol. 21(4/5): 521-525.
- Kuhn, J., 2007, Personal communication (phone call). Montana Department of Environmental Quality, Helena, Montana.
- Landau, M. and J.W. Tucker Jr. 1984. "Acute Toxicity of EDB and Aldicarb to Young of Two Estuarine Fish Species." Bull. Environ. Contam. Toxicol. 33(2): 127-132.
- Layton, D., and Daniels, J. 1999. Chapter 5: "Potential Impact of Ethanol-containing Gasoline on Surface Water Resources." in Cannon, G., ed., Health and Environmental Assessment of the Use of Ethanol as a Fuel Oxygenate: Volume 4: Potential Ground and Surface Water Impacts, University of California, Lawrence Livermore National Laboratory, UCRL-AR-135949
- Iott, T. 2001 "Effects of Ethanol on Aquatic Life." in Frye, E., ed., Health, Environmental, and Economic Impacts of Adding Ethanol to Gasoline in the Northeast States, Water Resources and Associated Health Impacts. Volume 3, pp. 49-53.
- LeBlanc. G.A. 1980. "Acute Toxicity of Priority Pollutants to Water Flea (Daphnia magna)." Bulletin

- of Environmental Contaminant Toxicology. 24(5): 684-691.
- Mayer, F.L. Jr. and M.R. Ellersieck. 1986. Manual of Acute Toxicity: Interpretation and Database for 410 Chemicals and 66 Species of Freshwater Animals. Resource Publication # 160. U.S. Dept. of Interior, Fish and Wildlife Service. Washington, D.C. 505 p.
- Meyers, R. 2006. US Senate Committee on Environment and Public Works Hearing Statements, The Impact of the Elimination of MTBE.
- National Toxicology Program (NTP), U.S. Department of Health and Human Services. 1985. NTP Technical Report on the Toxicology and Carcinogenesis Studies of 1,2 Dichlorobenzene (odichlorobenzene) in F344/N Rats and B6C3F<sub>1</sub> Mice (Gavage Studies). NTP TR 255, NIH Publication No. 86-2511. Research Triangle Park, NC.
- Neff, J.M., J.W. Anderson, B.A. Cox, R.B. Laughlin Jr., S.S. Rossi, and H.E. Tatum. 1976. "Effects of Petroleum on Survival, Respiration, and Growth of Marine Animals." Am. Inst. Biol. Sci.: 516-539.
- O'Connor, M. J. 2006. Memorandum. Virginia Petroleum, Convenience and Grocery Association, Webpage on MTBE Phaseout/Ethanol Introduction in Virginia RFG Markets Within 30 Days.
- Pearson, C.R. and G. McConnell. 1975. "Chlorinated C1 and C2 Hydrocarbons in the Marine Environment." Proc. R. Soc. Lond. B Biol. Sci. 1989: 305-332,
- Powers, S., Alvarez, P. J., and Rice, D.W., 2001, Chapter 1: "Increased Use of Ethanol in Gasoline and Potential Groundwater Impacts." in Rice, D. W. and Depue, R.T., ed. Environmental Assessment of the Use of Ethanol as a Fuel Oxygenate: Subsurface Fate and Transport of Gasoline Containing Ethanol, Report to the California State Water Resources Control Board. UCRL-AR-145380.
- Rice, S.D. and R.E. Thomas. 1989. "Effect of Pre-Treatment Exposures of Toluene or Naphthalene on the Tolerance of Pink Salmon (*Oncorhynchus gorbuscha*) and Kelp Shrimp (*Eualis suckleyi*)." Comp. Biochem. Physiol. C. 94(1): 289-293.
- RTDF Bioremediation Consortium. 1998. Natural Attenuation of Chlorinated Solvents in Groundwater.
- Thomas, D.H. and J.J. Delfino. 1991. "A Gas Chromatographic/ Chemical Indicator Approach to Assessing Groundwater Contamination by Petroleum Products." Groundwater Monitoring Review. Fall 1991. pp. 90-100
- Trucco, R.G., F.R. Engelhardt, and B. Stacey. 1983. "Toxicity, Accumulation and Clearance of Aromatic Hydrocarbons in Daphnia Pulex." Environ. Poll. Serv. A. 31(3): 191-202.
- Texas State Energy Conservation Office, 2007a, Webpage on Ethanol. <a href="http://www.seco.cpa.state.tx.us/re\_ethanol.htm">http://www.seco.cpa.state.tx.us/re\_ethanol.htm</a>
- ----, 2007b, Webpage on MTBE & Ethanol, http://www.seco.cpa.state.tx.us/re\_ethanol\_MTBE.htm
- Ulrich, G. 1999. The Fate and Transport of Ethanol-Blended Gasoline in the Environment: a Literature Review and Transport Modeling. Governors' Ethanol Coalition: Lincoln, Nebraska
- Walbridge, C.T., J.T. Fiandt, G.L. Phipps, and G.W. Holcombe. 1983. "Acute Toxicity of Ten Chlorinated Aliphatic Hydrocarbons to the Fathead Minnow (*Pimephales promelas*)." Arch. Enviorn. Contam. Toxicol. 12(6): 661-666.
- Ward, G.S., A.J. Tolmsoff, and S.R. Petrocelli. 1986. Acute Toxicity of Trichloroethylene to Saltwater Organisms. Bull. Environ. Contam. Toxicol. 37(6): 830-836.
- Yoshioka, Y., Y. Ose, and T. Sato. 1986. "Correlation of the Five Test Methods to Assess Chemical Toxicity and Relation to Physical Properties." Ecotoxicol. Environ. Saf. 12(2): 15-21.